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Aircraft study of the impact of lake-breeze circulations on trace gases and particles during BAQS-Met 2007

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Abstract

High time-resolved aircraft data, concurrent surface measurements and air quality model simulations were explored to diagnose the processes influencing aerosol chemistry under the influence of lake-breeze circulations in a polluted region of southwestern Ontario, Canada. The analysis was based upon horizontal aircraft transects at multiple altitudes across an entire lake-breeze circulation. Air mass boundaries due to lake-breeze fronts were identified in the aircraft meteorological and chemical data, which were consistent with the frontal locations determined from surface analyses. Observations and modelling support the interpretation of a lake-breeze circulation where pollutants were lofted at a lake-breeze front, transported in the synoptic flow, caught in a downdraft over the lake, and then confined by onshore flow. The detailed analysis led to the development of conceptual models that summarize the complex 3-D circulation patterns and their interaction with the synoptic flow. The identified air mass boundaries, the interpretation of the lake-breeze circulation, and best estimates for air parcel circulation times in the lake-breeze circulation (1.2 to 3.0 h) enabled formation rates of oxygenated organic aerosol (OOA/ ΔCO) and SO_4^{2-} to be determined. The formation rate for OOA, relative to excess CO, was found to be $2.5\text{--}6.2\ \mu\text{g m}^{-3}\ \text{ppmv}^{-1}\ \text{h}^{-1}$ and the SO_4^{2-} formation rate was $1.8\text{--}4.6\% \text{ h}^{-1}$. The formation rates are enhanced relative to regional background rates implying that lake-breeze circulations are an important dynamic in the formation of SO_4^{2-} and secondary organic aerosol. The presence of cumulus clouds associated with the lake-breeze fronts suggests that these enhancements could be due to cloud processes. Additionally, the effective confinement of pollutants along the shoreline may have limited pollutant dilution leading to elevated oxidant concentrations.

1 Introduction

The temporal and spatial structure of the boundary layer and lower troposphere near coastlines can be complex due to radiative property differences between land and water. This can be an important feature affecting the transport and transformation of air pollutants in such regions. Temperature gradients between cooler air over water and warmer air over land result in a pressure gradient that can initiate and sustain a lake (sea) breeze during the day and a land breeze at night. During the day, air over a lake moves inland in a shallow inflow layer (typically < 500 m) and air aloft over land moves offshore in a return flow to replace air from over the lake. As the cooler lake air moves over the warmer land surface, a thermal internal boundary layer (TIBL) is created which grows in height with an increase in inland distance (Lyons and Olsson, 1973; Garratt, 1990). At the leading edge of the lake breeze, air is forced upwards at the convergence zone (lake-breeze front) that separates the cooler lake air from the warmer air inland. The passage of a lake-breeze front is often characterized by increased upward motion, enhanced moisture and wind shear, decreased temperatures and changes in wind direction and speed (Lyons, 1972). Due to the upward motion of air at the lake-breeze front, a line of cumulus clouds may form along the frontal zone. At night, the temperature gradient is reversed and a land breeze is formed.

Numerous studies have identified the importance of lake (sea) breeze circulations to air quality in coastal urban areas (e.g. McElroy and Smith, 1986; Lu and Turco, 1994; McKendry et al., 1997; Cheng, 2002; Snyder and Strawbridge, 2004; Bouchlaghem et al., 2008; Lin et al., 2010). Specifically, a number of studies have been conducted in the Great Lakes area to examine lake breezes and their influence on air quality (e.g. Biggs and Graves, 1962; Lyons, 1972; Lyons and Cole, 1976; Reid et al., 1996; Sills 1998; Hastie et al., 1999). The region of southwestern Ontario often experiences poor air quality due to elevated O₃ and particulate matter (PM) concentrations (OME, 2008). The impact of the lakes and their associated lake-breeze circulations has been a complicating factor in elucidating chemical processes on a fine spatial scale in this

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region. For example, aircraft measurements showed that aged, polluted air masses over Lake Ontario were advected over land by lake breezes with abrupt increases in O₃ concentrations, as well as other pollutants such as NO_x and PAN (Reid et al., 1996; Hastie et al., 1999). However, the interaction between lake breezes, chemical transport and processes behind the increases is not well understood.

In the summer of 2007, the Border Air Quality and Meteorology Study (BAQS-Met) was conducted in southwestern Ontario to study the effects of lake breezes on air quality. This multi-agency, collaborative study was a unique opportunity to relate an intensive set of meteorological data to a comprehensive suite of trace gases and particle measurements in an area of frequent lake breeze activity. This paper presents a detailed analysis of a complete lake-breeze circulation using high spatially and temporally resolved meteorological and chemical measurements. Interpretation of the measurements in combination with air quality model simulations provides new insights into the complexities of lake-breeze circulations related to air mass processing.

2 Experimental design

The study area was located in the southern Great Lakes region between Lakes Huron, Erie and St. Clair (see Fig. 1). A meso-network of 50 stations spaced approximately 15 km apart collected measurements of meteorology, O₃ and PM concentrations. These measurements were made from 1 June to 31 August 2007. The sites relevant to the present analyses include Cottam, Essex, Woodslee, Lighthouse Cove, Lake St. Clair (LSC) Buoy and Sombra (Fig. 1). From 20 June to 10 July 2007, intensive measurements of gaseous compounds, particle chemistry and physics and meteorology were made at three supersites (Bear Creek (42.5359° N, -82.3892° W), Harrow (42.0330° N, -82.8933° W) and Ridgetown (42.4533° N, -81.8878° W) (Fig. 1), and from an aircraft and a ground-based mobile laboratory, CRUISER. All times are in local time (LT) i.e. Eastern Daylight Savings Time.

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The particle collection efficiency (CE) of the C-AMS for ambient measurements is typically determined through comparisons with other particle chemical measurements such as with the Particle Into Liquid Sampler-Ion Chromatography instrument (PILS-IC) and/or with mass estimated from instruments such as Scanning Mobility Particle Sizers (SMPS). The CE is a function of the particle transmission through the aerodynamic lens, the efficiency of particles being focussed by the lens and directed onto the vaporizer, and the extent to which particles bounce off the vaporizer. Many researchers have determined a CE ~ 0.5 for ambient particles (e.g. Dunlea et al., 2009; Kleinman et al., 2007; DeCarlo et al., 2008; Drewnick et al., 2004; Allan et al., 2004). However, higher CE values between 0.5 and 1 were found for particles dominated by ammonium nitrate and for acidic particles (Quinn et al., 2006; Kleinman et al., 2007). During BAQSMet, the aerosol neutralization ratio (ANR) defined as the molar ratio of ammonium to (sulphate + nitrate) indicated that the particles were neutralized most of the time with the exception of data taken in high concentration sulphate plumes, where the particles had not yet been neutralized by NH_3 . A comparison of mass concentrations from the C-AMS with those estimated from the PCASP indicated a CE of 0.5 for neutralized particles (ANR ~ 1.0), but approached 1.0 for acidic particles (ANR < 0.5). Therefore, the C-AMS data in this study were adjusted using CE = 0.5 for neutralized particles transitioning linearly to a CE = 1.0 for acidic particles. A pressure controlled inlet (PCI) was used in front of the AMS to remove variations in particle sizing and transmission due to pressure changes in the aerodynamic lens of the AMS (Jayne et al., 2000; Hayden et al., 2008; Bahreini et al., 2008). In the PCI, a low pressure region, between a 200 μm orifice upstream of the AMS inlet and a 130 μm orifice in the AMS inlet (replacing the standard 100 μm orifice) was maintained at a set point of 470 torr. This low pressure region was variably pumped so that the inlet pressure of the AMS was maintained at 1.3 torr. Under this configuration, no corrections to particle sizing were required for altitudes up to 3000 m a.g.l. during this study. Transmission efficiency experiments performed during and after the study indicated that particles were transmitted through the PCI with 100% efficiency.

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2.4 Model description

The Unified Regional Air-quality Modelling System (AURAMS) air quality model was used to support the BAQSMet field study. Details of the model are provided in Makar et al. (2010), but a brief description is presented here. The model has three main components: (1) a prognostic meteorological model GEM (Global Environmental Multiscale model: Côté et al., 1998); (2) an emissions processing system, (Sparse Matrix Operator Kernel Emissions: Houyoux et al., 2000; CEP, 2003); and (3) an off-line regional chemical transport model, the AURAMS Chemical Transport Model (CTM: cf. Cho et al., 2009; Makar et al., 2009; Gong et al., 2006). The AURAMS model was run with three nested grids at 42, 15 and 2.5 km horizontal resolution, driven by 15 km (for the two coarser resolutions) and 2.5 km GEM simulations. The driving meteorology simulations were created at the coarser (15 km) resolution from analyses updated every 6 h, with 12 h simulations having 6 h of discarded spinup, the final 6 h being stitched together for a continuous set of meteorological inputs for AURAMS. The driving meteorology was stored in 15 min timesteps for the coarser AURAMS simulations, and 2 min timesteps for the 2.5 km resolution simulation. Makar et al. (2010) provide a detailed evaluation of the model performance showing comparisons to observations with the aircraft and identified the presence of very local-scale features. Model output at 2.5 km horizontal resolution is used in conjunction with the measurements to evaluate pollutant sources and transport into the study region, and provides a more comprehensive analysis of pollutant behaviour and processing during a lake-breeze event.

In addition to the AURAMS simulations, high resolution forward and backward trajectories were calculated using the Canadian Meteorological Centre (CMC) trajectory model (D'Amours et al., 2001). This model made use of the 2.5 km resolution, 2 min timestep GEM wind fields that were also used to drive AURAMS at its highest resolution. Back trajectory endpoints and forward trajectory starting points were chosen to help elucidate the flow patterns associated with the driving meteorology, as will be discussed in more detail.

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3 Results and discussion

A brief summary of meteorological sampling conditions on aircraft flight days is provided in Table 1. A detailed analysis of the meteorology and lake breezes observed throughout the study has been described elsewhere (e.g. OME, 2008; Sills et al., 2011; Slowik et al., 2011). The synoptic wind flow was determined from the high resolution back trajectories computed for air parcels arriving into the study region at 500 m a.g.l. Large scale synoptic flow varied considerably, and lake breezes were identified on each day (Sills et al., 2011). The extent to which the lake-breeze circulations were deformed by the synoptic wind was used to classify lake breezes as Low Deformation (LD), Moderate Deformation (MD) or High Deformation (HD) (Sills et al., 2011).

The highest O₃ (max = 105 ppbv, mean = 73.0 ppbv) and SO₄²⁻ (max = 29.1 µg m⁻³, mean = 29.07 µg m⁻³) levels were observed from the aircraft on 25 June under light to moderate southwesterly synoptic flow that contributed to a poor air quality episode between 24–27 June. On 25 June, lake breezes developed in association with each of the lakes in the region and aircraft flights (Flights 4 and 5) were designed to sample across an entire LSC lake-breeze circulation at multiple altitudes. Lake-breeze fronts are shown as magenta lines in Fig. 1. The lake-breeze circulations on this day were typical of LD situations that occurred several times throughout the study period (Sills et al., 2011). In this paper, a complete lake-breeze circulation on 25 June is analyzed. Hourly surface mesoscale analyses (see Sills et al., 2011) are used as a guide to identify signatures of lake-breeze fronts in the aircraft meteorological data and compared with spatial and vertical changes in pollutant concentrations. Cross-sectional plots constructed from the aircraft data are analyzed in conjunction with model simulations to interpret pollutant transport and processing within the lake-breeze circulations. Conceptual models of the lake-breeze circulation are proposed followed by estimates of the lake breeze impacts on air mass processing of PM.

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3.1 Air mass history

As shown in Fig. 1, the study region is impacted by local, anthropogenic emissions from urban centres including Sarnia to the north, Detroit/Windsor in the west/northwest, Toledo to the southwest and Cleveland on the south shore of Lake Erie (LE). Emissions of SO₂ and NO_x from power generating plants and oil/chemical refining plants in the region include the St. Clair (USA) and Lambton (Canada) stations (north of Lake St. Clair [LSC]), Monroe (western shore of LE), and Avon Lake and Cleveland Electric (south shore of LE). Major SO₂ point sources (SO₂ emissions > 10 000 tons yr⁻¹, United States Environmental Protection Agency (EPA, <http://epa.gov/airmarkets>); Canadian National Pollutant Release Inventory (NPRI, www.ec.gc.ca/inrp-npri/default.asp?lan=en)) are shown as red circles in Fig. 1. The region is further affected by local mobile and agricultural precursor emissions, ship emissions along the St. Lawrence Seaway from LE and along the Detroit River through to Detroit, and by longer-range transport of pollutants from midwestern US states.

On 25 June 2007, a weak pressure ridge extended from the New England states in the United States across the Great Lakes. The synoptic flow was light from the southwest. In Fig. 2, high resolution back trajectories (D'Amours et al., 2001) are shown arriving into the study region during the Flight 4 time period (11:01–13:24 LT). The arrival points correspond to aircraft locations, times and altitudes selected along the Flight 4 track. Trajectories are coloured as a function of altitude. Figure 2 indicates that the predominant flow on 25 June was from the south to west-southwest with air masses having travelled over urban areas and power plants at the west end of LE. Figure 2 also shows that the back trajectories are sensitive to perturbations in air mass flow due to lake breezes. For example, over the Harrow site, the upwind flow was westerly, with transport of emissions from the west end of Lake Erie, but after the onset of the lake breezes (11:00 LT), the back trajectories switched to southerly closer to the site. This is consistent with surface measurements at Harrow that showed a change in wind direction from westerly to southerly at this time. The back trajectory at the south

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end of LSC also shows a change in direction from westerly to northeasterly in response to the LSC onshore lake breeze.

3.2 Lake-breeze identification

Mesoscale analyses for each hour (Sills et al., 2011) and the inland progression of the lake breezes from 12:00 to 17:00 LT p.m. on 25 June 2007 is presented in Fig. 1. The positions of the lake-breeze fronts are shown as magenta lines. Lake breezes on this day were first detected at 11:00 LT and persisted until 21:00 LT. At 12:00 LT (Fig. 1a), lake-breeze fronts were detected along the south and north shores of LE, as well as almost all the way around LSC. The north shore LE front progressed northward toward LSC until the LE and LSC fronts merged along the south shore of LSC at 15:00 LT (Fig. 1d). The merged fronts then remained quasi-stationary until after 17:00 LT.

On 25 June 2007, two of the three aircraft flights sampled across the LSC lake-breeze circulation. Aircraft tracks for Flights 4 and 5 are shown in Fig. 1a and f respectively, for the corresponding time interval in which they occurred. Flight 4 included horizontal transects between the north shore of LE (near the Harrow supersite) and near Lambton (north of the Sombra site) at 1560 m a.g.l. (free troposphere), 800 m a.g.l. and 300 m a.g.l. Flight 5 included a similar flight pattern with multiple transects at altitudes of 2600 m a.g.l., 1560 m a.g.l., 800 m a.g.l., 460 m a.g.l. and 300 m a.g.l. At the beginning of Flight 5, the aircraft performed a vertical spiral over the LSC buoy extending from 200–2600 m a.g.l. Using the analyzed surface lake-breeze front locations as a guide, the aircraft meteorological and chemical data were examined to identify signatures due to lake breezes. On 23 June, low deformation lake breezes also occurred with even stronger lake-breeze convergence zones (Sills et al., 2011). Although, pollutant concentrations were low during this period of northwesterly synoptic flow (Table 2), aircraft data on this day were similarly analyzed for lake-breeze boundaries. It was found that on June 23 sharp changes in the dewpoint temperature and sustained upward gust velocities were good indicators of lake-breeze front crossings. This provided confidence that using these parameters to identify lake-breeze fronts on 25 June was applicable.

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This approach was also consistent with an independent detailed analysis of all flight data.

3.3 Aircraft observations of trace gases, particles and meteorology

The wind direction, dewpoint temperature, vertical gust velocity, CO, SO₂, particulate SO₄²⁻, and particulate organics (OA) are shown for Flights 4 and 5 in Figs. 3 and 4, respectively. The bottom axis is the distance between Lake Erie (LE) and Lambton (just past Sombra) (see Fig. 2 for locations), the blue horizontal bar represents LSC and the grey arrow indicates the aircraft flight direction. The blue and green arrows indicate the position of the lake-breeze fronts as identified from the aircraft data. The light blue boxes are selected time slices that are discussed in Sect. 3.7. Although the fronts above the surface were likely not in the same horizontal location as at the surface because of sloped frontal zones, the difference was likely small, i.e. <1 km. Thus, the mesoscale surface analyses were used to guide the front locations in the aircraft data. An additional complication considered in the analysis was the movement of the fronts over time i.e. between the 800 and 300 m a.g.l. aircraft passes, the LE front migrated northward approximately 10 km. In Flight 4, the 1560 m a.g.l. track was in the free troposphere above an estimated 1200 m a.g.l. synoptic inversion height (Fig. 3a1, a2). CO mixing ratios were <140 ppbv and SO₂ was below detectable levels. Particle mass concentrations of OA and SO₄²⁻ ranged between 0.5–3 μg m⁻³. These observations are indicative of homogeneous regional background air. In the 800 m a.g.l. transect, travelling from LE to Lambton (Fig. 3b1), the LE lake-breeze front can be clearly identified by a 2 °C increase in the dewpoint from 18 °C to 20 °C (blue arrow at 11:56 LT), which coincides with the position of the lake-breeze front based on analyses of surface meteorological data. The increase in the dewpoint is due to updrafts at the front bringing up moister air from the surface. There was also a sudden increase in the aircraft vertical gusts (Fig. 3b1) and a visible line of clouds. At this altitude the aircraft was below the synoptic inversion, but above a shallower LE lake-breeze inflow layer (on

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the lake side of the front), and thus, the relatively low chemical levels (Fig. 3b2) reflect “residual” air or remnants of the previous day’s convective boundary layer (Sills et al., 2011). The aircraft then crossed the LE front and measurements showed an increase in primary and secondary pollutants consistent with a polluted convective boundary layer. Elevated pollutant concentrations were observed over land from the LE lake-breeze front to the south shore of LSC. Evidence of the south shore LSC lake-breeze front at 800 m a.g.l. was more difficult to identify based on meteorology alone. Since CO is a relatively long-lived species in the troposphere, these measurements were examined for perturbations related to lake-breeze boundaries. Guided by the location of the surface lake-breeze front, a brief, sharp change in CO of 24 ppbv (green solid arrow at 12:01 LT), was observed (Fig. 3b2). This was coincident with an increase in the vertical gusts (Fig. 3b1). On the LSC side of the LSC front, pollutants remained elevated followed by a sharp decrease over the south shore of LSC. The identification of the lake-breeze front north of LSC (green dashed arrow) is based on a change in the dewpoint from 21° to 19°C (12:12 LT); there were no discernible features suggesting the presence of a surface-based front.

At Lambton, the aircraft turned around and travelled southward toward Harrow at 300 m a.g.l. A LSC front north of LSC was not detectable in the aircraft data (Fig. 3c1, c2). As the aircraft passed over the south shore of LSC at 12:35 LT, the wind direction changed from southerly to northeasterly (Fig. 3c1). A northerly wind direction at this time was also measured at the LSC buoy (42.425° N, -82.558° W), Woodslee (42.212° N, -82.748° W) and Lighthouse Cove (42.292° N, -82.522° W) mesonet stations (see Fig. 1 for site locations). This is due to the onshore LSC lake breeze and demonstrates that the aircraft was successful in penetrating and sampling in the shallow lake-breeze inflow layer at 300 m a.g.l. Highest pollutant concentrations were between LE and just over the south shore of LSC which were generally consistent with levels at 800 m a.g.l. suggesting the development of a well-mixed convective boundary layer (CBL) (Fig. 3c2). Chan et al. (2011) also reported high surface concentrations of CO, SO₂ and SO₄²⁻ near this time at the LSC shoreline to the northeast of the

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flight path. Concentrations were also observed to decrease sharply over the south end of LSC. Sharp changes in the dewpoint of ~1°C signified the presence of the LE (12:39 LT) and LSC (12:37 LT) lake-breeze fronts, as supported by the surface mesoanalyses. The LSC front was also identified in the aircraft data by the onshore wind direction. Between Harrow and the LE lake-breeze front, plumes of SO₂ and SO₄²⁻ were observed having been advected from over LE.

During the time between Flight 4 and Flight 5, the LE lake-breeze front pushed north ~35 km and merged with the LSC front along the south shore of LSC by 15:00 (Fig. 1d). As a result, the CBL air between LE and LSC was replaced with lake air leaving the study region being largely influenced by lake-breeze circulations and lake-modified air masses. The first two transects in Flight 5 at 2600 m a.g.l. and 1560 m a.g.l. were in free tropospheric air. Figure 4a1, a2 shows the measurements from the transect at 1560 m a.g.l. indicating that the predominant flow was still from the southwest. As in Flight 4, low pollutant concentrations at the higher altitudes are believed to reflect regional background air. The measurements in the remaining transects (Fig. 4b–d) reflect residual air influenced or modified by lake-breeze return flow. Multiple, positive peaks in the updraft velocity at the three altitudes within the residual layer (16:36 LT (800 m a.g.l.), 16:43 LT (460 m a.g.l.), 17:20 LT (300 m a.g.l.)) are interpreted as updrafts along the merged LE/LSC lake-breeze front. Pollutant concentrations in this later afternoon flight were generally lower and less variable compared to Flight 4, but elevated pollutant concentrations were observed north of LSC, possibly from the CBL between LE and LSC, or pollutants from over LSC earlier that were advected northward and then mixed upward. North of LSC, SO₄²⁻ was >15 µg m⁻³ and CO, in particular, was variable ranging between 180–390 ppbv indicating a non-homogenous air mass. As the aircraft went further north just past Sombra, the measurements were likely influenced by power plant emissions from the Lambton area that were advected northeast in the southwesterly flow.

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3.4 Distribution and evolution of trace gases and particles

Data from the aircraft transects, averaged to 10 s, and measurements from eight surface sites were interpolated using an ordinary linear kriging method (GS + ver. 5.1.1, Gamma Design Software) to produce cross-sectional plots of CO, SO₂ and vertical winds (Figs. 5, 6). Overlaid on the plots are the interpolated plan-view horizontal wind directions (along with the corresponding orientation of the north arrow) and the size of the arrows represents the magnitude of the wind speed. The vertical wind component is not included in the wind direction, but is shown separately in the third panel of the figures to highlight the measured vertical motions. It was assumed that physical and chemical processes were sufficiently stable to build 2D cross sections from the aircraft transects. CO and SO₂ were chosen because they are both good tracers for different emission sources, both had fast instrument time responses (Table 2) and due to its 1 s time-response, CO appeared to be a good chemical tracer for lake-breeze motions. The surface sites chosen for this analysis were within 12.4 km horizontal distance of the Flight 4 and 5 mean flight path and included Harrow (3.3 km), Cottam (1.3 km), Essex (5.7 km), Woodslee (0 km), Lighthouse Cove (12.4 km), the LSC Buoy (3.8 km), Bear Creek (11.4 km) and Sombra (0.8 km) (Fig. 1a, f). The sites are indicated as black circles at the bottom of Figs. 5 and 6.

In Fig. 5, the LE lake-breeze front separates LE air and polluted CBL air. In Fig. 5c, a return flow associated with this front is not present, but an area of reduced wind speed (red circle) suggests that it may have been masked by the larger synoptic flow with a resulting net flow along the axis of the transect (southwesterly). Updrafts associated with convective motion were observed in the CBL, and a significant downdraft was measured on the LSC side of the LSC lake-breeze front. The synoptic inversion is estimated to be ~1200 m a.g.l. In Fig. 5, CO and SO₂ were observed to be well mixed vertically within the CBL consistent with increased convection over the land surface. Transport of emissions from Detroit/Windsor and western LE in the southwesterly flow likely contributed to the elevated pollutant concentrations observed over land between

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LE and LSC. A horizontal gradient in CO mixing ratios was observed with highest levels further north compared to SO₂ mixing ratios which were elevated in the entire region between LE and LSC. This is consistent with the back trajectories (Fig. 2) that show air masses that arrived near Harrow had crossed over SO₂ emission sources at the west end of LE, whereas other trajectories arriving at points further north of Harrow also crossed over the Detroit/Windsor urban area. At the Harrow supersite, SO₂ was observed to be “plume-like”; elevated SO₂ was observed between 10:00–11:20 LT with mixing ratios reaching >100 ppb. Although lake-breeze circulations were present around LE with onshore flow near Toledo, in order for plumes of SO₂ to travel northward, they were either lofted over top of the lake-breeze inflow and/or were in transit prior to the formation of the lake breezes. Based on aircraft measurements, the spatial extent of the SO₂ plumes is estimated to range between 7.0–13.5 km horizontally; this indicates that the plumes were remarkably intact suggesting minimal dispersion, and thus mostly transported over the lake. Lyons and Pease (1973) found that plumes from power plants travelled long distances over Lake Michigan with little dilution and attributed this to the absence of convective mixing over the lakes. It might be expected that if the plumes had travelled over top of the lake-breeze inflow, they would have undergone some mixing and dilution. The back trajectories indicate a 4–6 h transit time from the west end of LE to arrive near Harrow (Fig. 2) and since this is prior to the onset of the lake breezes, the likely explanation for the SO₂ was that the plumes were already in transit and not affected by local lake breeze mixing at the western end of LE. In addition, the observations from the later afternoon flight do not show an influence from SO₂ emissions (except north of LSC from other sources, Figs. 4, 6) suggesting that the lake-breeze onshore flow near Toledo, which was still present at this time (Fig. 1), impedes transport to the study region even though the large scale flow continues to be southwesterly.

Elevated mixing ratios of CO and SO₂ were also observed on the LSC side of the LSC lake-breeze front (Figs. 3, 5). The sharp decrease of pollutant levels at the LSC shore points toward important differences between lake and land surfaces and

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influences from the lake-breeze circulation. It is hypothesized that pollutants in the CBL were lofted upward at the LSC lake-breeze front, transported northward in the synoptic flow, transported in the downdraft on the north side the front, and then confined by the LSC onshore flow along the south shore of LSC. This is plausible since
5 some of the CBL air closest to the LSC front would be expected to be advected in the upward vertical motion at the frontal convergence zone, and the downdraft that was measured behind the front (Fig. 5c) would draw air from above. This would result in the transport, at least partially, of air from the CBL by the downdraft behind the LSC front. The pollutants are then confined along the LSC shoreline by the onshore lake
10 breeze. The development of a thermal internal boundary layer (TIBL) over land behind the lake-breeze front would limit vertical mixing; as the TIBL grows vertically with inland distance it does not get as high as the fully developed CBL in the short distance between the lake shore and the front. Figure 3b2, c2 shows that CO and SO₂ behind the LSC front are less vertically developed than in the CBL supporting this theory.
15 The lake-breeze effect of confining pollutants is also evident at the surface. During the time of Flight 4, significant levels of SO₂ (> 50 ppb) and SO₄²⁻ (>35 µg m⁻³) were observed by the CRUISER mobile laboratory from 12:40–13:10 LT at Mitchell's Bay, located on the shoreline at the eastern edge of LSC (see Fig. 1) (Chan et al., 2011). Also along the eastern shoreline, at Bear Creek, SO₂ was 15–20 ppb from 11:00–12:30 LT.
20 There was no evidence of any significant pollutant buildup north of LSC, attributed to the weaker offshore lake-breeze circulation compared to the south and east sides of the lake. Mechanisms to transport pollutants within a lake-breeze circulation have been identified in previous studies (Lyons, 1972; Lyons and Cole, 1976; Sills et al., 1998) and are known to be effective in confining pollutants to coastal regions (Simpson, 1994; Lu and Turco, 1995). Dynamics at the lake-breeze front have been studied to understand
25 effects on pollutants. Hastie et al. (1999) noted that the arrival of the lake breeze coincided with maximum trace gas concentrations with rapid decreases behind the front suggesting that vertical mixing generated by the front could result in entrainment of cleaner air from above the inflow layer. Kitada and Kitagawa (1990) showed that the

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vertical profile of pollutants was strongly impacted by the micro-scale features of the sea breeze such as the transport of pollutants in the downdraft behind the front and suggested that the most aged air mass would be found in the upper part of the circulation behind the front. Since lake and land breezes are quasi-closed circulations and
5 pollutants emitted into them can be recirculated (Lyons, 1972), it might be expected that air in the downdraft of the LSC lake-breeze front would exhibit characteristics of a more aged air mass relative to the air in the CBL. The impact of lake breezes on air mass processing is discussed further in Sect. 3.7.

In Fig. 6, a signature of the merged LE/LSC front is not apparent, but surface analyses show that the front was still present just offshore at 17:00 LT and by 19:00 LT
10 the LSC lake breeze moved back onshore, with a shift to onshore winds and a rapid increase in dewpoint. The height of the synoptic inversion, determined from a vertical profile performed over the LSC buoy (15:47–15:55 LT), was 1300 m a.g.l. It is noted that strong east-southeast winds were measured above the merged front at the
15 2600 m a.g.l. level which is consistent with 08:00 LT and 20:00 LT DTX soundings showing winds with an easterly component between 2 and 3 km. In Flight 5, pollutant concentrations, compared to Flight 4, were more homogeneous across the flight transect. In Fig. 6, with the exception of the north end of the flight track near Lambton, CO and SO₂ levels were lower and less variable compared to Flight 4. The region is largely
20 enveloped in lake-modified air and pollutants in the CBL observed in Flight 4 appear to have been displaced by the merging LE/LSC lake-breeze fronts. Thus, by the late afternoon, only low pollutant concentrations remained between LE and LSC. The region having been being fed by regional lake air from the south is expected to exhibit characteristics of a more aged air mass and is discussed further in Sect. 3.7. It is speculated
25 that pollutants in the CBL between LE and LSC in Flight 4 were advected upward and carried northward by the southerly flow and/or a LSC return flow resulting in increased concentrations of pollutants northeast of LSC. Pollutants would be transported to the surface by land-based convection.

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3.5 Analysis of model-generated lake-breeze circulation

A more comprehensive view of the lake-breeze circulations is possible through analysis of model data. In Fig. 7, model-predicted SO₂ mixing ratios overlaid with wind direction are shown at 12:00 LT for altitudes of (a) 285 m a.g.l. and (b) 815 m a.g.l. The model altitudes are the closest available to the 300 and 800 m a.g.l. aircraft transects, and the model simulation time corresponds to the middle of Flight 4. After 15:00 LT, the model predicts winds shifting to come from the southeast, while the observations indicate a continuation of southwesterly flow (Figs. 4 and 6). As such, this discussion is reserved to that of the Flight 4 time period because the model in the later afternoon is not simulating the synoptic wind direction adequately. Figure 7a shows two zones of convergence (solid magenta lines) corresponding to the model's predicted locations of the LE and LSC fronts across the Flight 4 aircraft transect (red dashed line). Figure 7b shows a region of divergence (dashed magenta lines) along the north shore of LE in response to the convergence at lower levels. The modelled wind direction along the aircraft transect is quite similar to the measured winds shown in Fig. 3b1 and c1 for the corresponding altitudes. This provides confidence that the model is correctly simulating both the synoptic flow and lake-breeze circulation at this time. Figure 7 shows that high concentrations of SO₂ are located over LE (and over the Lambton area to the north-east of LSC); these pollutants originated from sources on the west end of the lake and were advected into the study region by 12:00 LT, the Flight 4 time period.

In Fig. 8, vertical cross-sections of model-predicted (a) CO, (b) SO₂ and (c) SO₄²⁻ are shown along the aircraft flight track at 12:00 LT (red dashed line in Fig. 7) corresponding to the Flight 4 time period. The wind direction overlaid is the 3-D wind field in the plane of the cross-section and, thus can be used to indicate horizontal and vertical motion in the plane of the cross-section. Consistent with the observations, the wind vectors show convergence zones marking the LE (blue arrow) and LSC lake-breeze (green arrow) fronts at lower altitudes (<~600 m a.g.l.). The model results suggest a circulation where air south of the LSC lake-breeze front moves upwards and north-westwards from the

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southeastern LSC surface towards Detroit. Then air moves east-northeast back over LSC at ~800 m a.g.l., moves downward in the downdraft over LSC, moves ashore in the lake breeze and returns to Detroit at ~300 m a.g.l.; the black arrows in Fig. 8a depicts this motion in the plane of the flight path. The model shows a similar distribution in pollutants along the flight track as in Figs. 3 and 5. High pollutant levels are predicted to be south of LSC with CO and SO₂ well-mixed up to 1500 m a.g.l., consistent with the observations. However, the model shows elevated CO mixing ratios in a narrow region at the LE lake-breeze front. Also, the model has high levels of CO, SO₂ and SO₄²⁻ at altitudes above 600 m a.g.l. over LSC, absent in the observations. In the CBL between the LE and LSC fronts, CO mixing ratios of 160–170 ppbv are lower than the observations (180–340 ppbv). At the LSC front the model simulates a sharp decrease in all three species on the lake side of the LSC front, whereas the observations show higher values extending closer to the shore. In the circulation behind the LSC front, the model shows the downward transport of pollutants from aloft over the south end of LSC and the LSC inflow appears to move the pollutants along the south shore of LSC consistent with the interpretation of the measurements.

To further investigate the LSC lake-breeze circulation, forward trajectories shown in Fig. 9 were computed starting at several locations along the LSC shoreline within the lake-breeze inflow at 12:00 LT and 225 m a.g.l. The trajectories, derived from the same GEM meteorological data used in the air quality simulations, show a circulation along the LSC lake-breeze front that forms a tighter, nearly complete circuit at the easternmost shoreline departure point (Fig. 9a) with a recirculation time of 2 h (12:00–14:00 LT). Moving toward Detroit (Fig. 9b, c), the trajectories become more helical in nature. For example, the trajectory for the 12:00 LT departure point depicted in Fig. 9b starts at the surface, rapidly climbs to altitudes of 1000 m a.g.l. two hours later, then falls over LSC and continues to Detroit close to the surface. Figure 9d shows a forward trajectory starting at 08:00 LT at 225 m a.g.l. over Windsor, with subsequent motion to the east-northeast at altitudes of 1000 to 700 m a.g.l., subsidence over LSC (12:00 LT), then subsequent helical recirculation back to the west, along the south shore of the lake.

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the atmosphere. Both OOA and SO_4^{2-} are secondary aerosols resulting from concurrent chemical processing in an air parcel, and as such they are expected to be related in air masses on regional scales (i.e., at locations at least a certain distance downwind ($\sim >100$ km) of the main locations of their precursor emissions). However, the relative magnitudes of OOA and SO_4^{2-} formation depend on the mixture and concentrations of SO_2 and precursor hydrocarbons. Figure 12 shows that the relationship between $\text{OOA}/\Delta\text{CO}$ and $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ for Flight 5 is confined to a range of higher OOA production ($200 \mu\text{g m}^{-3} \text{ppmv}^{-1} \%^{-1}$, upper limit), whereas in Flight 4, a range extending to lower OOA production ($80 \mu\text{g m}^{-3} \text{ppmv}^{-1} \%^{-1}$), is observed, relative to SO_4^{2-} production. The relationship between $\text{OOA}/\Delta\text{CO}$ and $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ provides a relatively convenient gauge of how much and how fast OOA is formed since SO_2 conversion in clear air is a better understood process and can be obtained through modelling. The rate of SO_2 conversion in clear air has been reported at $1\text{--}3\% \text{h}^{-1}$ (Newman et al., 1981; Davis et al., 1979; Luria and Sievering, 1991) and can be further derived from the model. The modelled SO_4^{2-} production rate at 12:00 LT for 285 m a.g.l. for the study domain is shown in Fig. 13 and indicates that the regional background SO_4^{2-} production rate (non-plume influenced) is about $1\text{--}2\% \text{h}^{-1}$, though much higher ($>20\% \text{h}^{-1}$) closer to sources and high concentration plumes. The OOA formation rate for regional background air is determined by multiplying the slope in Fig. 12 (Flight 4, $m = 71.39 \mu\text{g m}^{-3} \text{ppmv}^{-1} \cdot \text{fraction}$) by the regional background SO_4^{2-} formation rates taken from Fig. 13 ($1\text{--}2\% \text{h}^{-1}$). Thus, the regional background OOA formation rate is $0.71\text{--}1.4 \mu\text{g m}^{-3} \text{ppmv}^{-1} \text{h}^{-1}$. The regional background formation rates of OOA and SO_4^{2-} are used as reference points in the following analysis for determining if the production rates could be enhanced in the lake-breeze recirculation.

Following on the analysis of air mass boundaries and movement in response to the lake breezes (summarized in Fig. 10), time slices of aircraft data were selected that represent different air mass “types” (shown as shaded boxes in Figs. 3 and 4). The selection of air mass types is as follows: RegB represents regional air along flight tracks

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above the boundary layer; LE represents air originating over LE and characterized by low pollutant concentrations with the exception of SO_2 and SO_4^{2-} from power plants across LE; CBL is a section of the boundary layer closest to the LSC lake-breeze front and most influenced by lake-breeze circulations; AF represents air in the downdraft region of the LSC lake-breeze front; and LSC is air least affected by recent land-based emissions and lake-breeze circulations. Changes in the chemical characteristics of OOA and SO_4^{2-} in these air masses are expected to reflect the influence of the lake-breeze circulations.

In Fig. 14, the average and standard deviation of the $\text{OOA}/\Delta\text{CO}$ and $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ ratios are shown for each air mass type. The $\text{OOA}/\Delta\text{CO}$ ratios in Flight 4 span a larger range of values, $81.2 \mu\text{g m}^{-3} \text{ppmv}^{-1}$, compared to Flight 5 ($26.1 \mu\text{g m}^{-3} \text{ppmv}^{-1}$). Similarly, the $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ ratio is more variable in Flight 4 compared to Flight 5. In addition, the $\text{OOA}/\Delta\text{CO}$ ratio shows a general increase from Flight 4 to Flight 5 of $10.9\text{--}24.2 \mu\text{g m}^{-3} \text{ppmv}^{-1}$. These ratios highlight the spatial heterogeneity in the earlier flight, largely induced by the separation of air mass types due to the lake-breeze boundaries, and the subsequent transition to a comparatively homogeneous air mass during the later afternoon flight. The highest $\text{OOA}/\Delta\text{CO}$ ratios in Flight 4 are in RegB, comparable to RegB values in Flight 5, suggesting similarities in the chemical composition of regional background air arriving in the study region throughout the day. In Flight 4 (Fig. 14a), the aircraft measurements in the CBL reflect a mixture of pollutant sources from Detroit/Windsor and from the west end of LE, and as such the low $\text{OOA}/\Delta\text{CO}$ value, $20.7 \pm 6.0 \mu\text{g m}^{-3} \text{ppmv}^{-1}$ indicates limited photochemical processing. This is consistent with reported values for unprocessed urban air. DeCarlo et al. (2008) note that the lowest ratio for relatively “fresh” urban air over Mexico City and surrounding region was $\sim 35 \mu\text{g s m}^{-3} \text{ppm}^{-1}$. For similarly fresh urban air, Kleinman et al. (2008) showed a ratio of $\sim 18 \mu\text{g m}^{-3} \text{ppmv}^{-1}$. Figure 14b shows $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ ratios in the CBL that are consistent with an air mass that has undergone limited processing of SO_2 . Compared to the CBL, the ratios over the lakes (LE and LSC) in Flight 4 reflect air masses that have been further processed.

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In Flight 4, as most of these air masses are “unconnected” due to the lake-breeze induced boundaries, the variability in the extent of processing is not surprising. However, based on the interpretation of the LSC recirculation inferred in Flight 4 (Figs. 10a, 11), air masses in the CBL appear to be connected to that behind the LSC lake-breeze front (AF). Particles in the CBL are hypothesized to move aloft over the LSC front towards the north into the zone labelled as AF and then to descend and possibly be recirculated through the onshore flow. Thus, the AF and CBL particles can be compared to determine if the lake-breeze recirculation results in changes in particle mass locally. This requires an estimate of the processing times associated with this circulation which can be derived from aircraft observations, model data, and through forward trajectory analyses. The aircraft measurements show a 2.0 m s^{-1} wind speed within the LSC inflow layer and the return flow at 800 m a.g.l. was 2.8 m s^{-1} , likely because of additional synoptic flow at 800 m a.g.l. From Fig. 5, the distance from the LSC front to the downdraft region over LSC is estimated to be 4.2–12.0 km and if the circuit distance travelled by an air parcel was along the aircraft transect, this distance would double to 8.4–24.0 km. Including a vertical distance component of $0.8 \text{ km} \times 2 = 1.6 \text{ km}$ results in a total distance of 10.0–25.6 km. Using appropriate wind speeds, this distance translates into a total circuit time of 1.2–3.1 h. For the model output in Fig. 8, wind speeds were 2.5 m s^{-1} in the inflow and return flows with a total distance travelled between 14.4–25.6 km (depending on how far over LSC the recirculation pathway might extend) and this distance translates into a total circuit time of 1.6–2.8 h, consistent with that derived based on measurement data. Forward trajectories shown in Fig. 9 indicate that a recirculation time of 2–3 h is reasonable. With these three methods, the recirculation time at the LSC front is somewhere between 1.2–3.0 h. By comparison, using tetrons released into lake-breeze circulations, Lyons and Olsson (1973) showed complete lake-breeze cycles ranging from 1.5–2.0 h.

In Fig. 14a (Flight 4), the OOA/ ΔCO ratio in the AF air mass (behind the LSC lake-breeze front) ($28.1 \pm 2.8 \mu\text{g m}^{-3} \text{ ppmv}^{-1}$) is higher compared to the CBL by $7.4 \mu\text{g m}^{-3} \text{ ppmv}^{-1}$. If the complete recirculation time is 1.2 h, the rate of increase

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in the OOA/ ΔCO ratio is $6.2 \mu\text{g m}^{-3} \text{ ppmv}^{-1} \text{ h}^{-1}$, but if the recirculation time is closer to 3.0 h, the increase corresponds to $2.5 \mu\text{g m}^{-3} \text{ ppmv}^{-1} \text{ h}^{-1}$. Thus, the OOA formation rate, relative to excess CO, determined in the lake-breeze recirculation, $2.5\text{--}6.2 \mu\text{g m}^{-3} \text{ ppmv}^{-1} \text{ h}^{-1}$, shows an enhancement over the regional background formation values inferred above ($0.71\text{--}1.4 \mu\text{g m}^{-3} \text{ ppmv}^{-1} \text{ h}^{-1}$). From Fig. 14b, the SO_2 conversion rate can be similarly estimated. The average $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ ratio in the AF air mass of 0.24 ± 0.03 and in the CBL of 0.24 ± 0.03 results in an estimated SO_2 conversion rate to range between $1.8\% \text{ h}^{-1}$ (3 h recirculation) and $4.6\% \text{ h}^{-1}$ (1.2 h recirculation). This suggests an enhancement for the shorter estimated recirculation times over the non-plume influenced regional background formation rate ($1\text{--}2\% \text{ h}^{-1}$) as predicted by the model. Enhancements in SO_4^{2-} and OOA formation rates, relative to regional background formation rates, implies that lake-breeze circulations are an important dynamic in the formation of SO_4^{2-} and SOA.

One possible explanation for such enhancements is through cloud processing. Cumulus clouds were in fact observed all along the LSC front and in the CBL between the LE and LSC fronts, especially between 12:00 and 14:00 LT. The presence of such clouds can significantly increase the rate of SO_2 oxidation (Luria and Sievering, 1991; Leaitch, 1996; Joos and Baltensperger, 1991) and SOA formation (Lim et al., 2005; Yu et al., 2005; Blando and Turpin, 2000). A second explanation is that the effective confinement of pollutants behind the LSC front may have prevented the dilution of pollutants from further mixing and led to elevated oxidant concentrations, and therefore enhanced oxidation of primary pollutants including SO_2 and organics. Both processes may be at work in the study region. Regardless of the processes, the lake-breeze front apparently led to locally enhanced SO_2 (and probably organic aerosol precursors) and increased production of SO_4^{2-} and SOA.

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4 Conclusions

High time-resolved aircraft data, concurrent surface measurements and meteorological and air quality model simulations were explored in a highly integrated manner to diagnose the processes influencing aerosol chemistry for a polluted BAQS-Met case associated with a well-developed lake-breeze circulation. This was based upon two aircraft flights (Flights 4 and 5 on 25 June 2007) in southwestern Ontario where horizontal transects across the entire lake-breeze circulation at multiple altitudes were performed. Air mass boundaries due to lake-breeze fronts were identified in the aircraft meteorological (dewpoint temperature and vertical gust velocity) and chemical data, which were consistent with the frontal locations determined from surface observations and satellite and radar images (Sills et al., 2011). The meteorological model was also found to simulate the conditions during the first flight reasonably well.

Cross-sectional plots created from the aircraft horizontal transects indicated that in the early afternoon flight (Flight 4), elevated mixing ratios of CO and SO₂ were well-mixed vertically within a convective boundary layer (CBL) and bounded by the Lake Erie (LE) and Lake St. Clair (LSC) fronts. Back trajectories indicated that the origin of these pollutants was from the west end of LE and the Detroit/Windsor area. Elevated pollutant concentrations were also observed aloft on the lake side of the LSC front and also closer to the surface, which was assumed to be due to subsiding air or downdrafts both observed and modelled. This area of high concentration did not extend far away from the front, thus indicating that the circulation pattern tended to confine pollutants in the lake-modified air relatively close to the front. By the late afternoon (Flight 5), the region was largely enveloped in lake air and pollutants observed in Flight 4 were displaced by the merging LE and LSC fronts. The hypothesized fate of these pollutants was transport aloft toward the north where they were convectively mixed to the surface north of LSC. This is supported by observations and modelling during Flight 4 showing that pollutants in the CBL were lofted in the upward vertical motions at the LE and LSC lake-breeze fronts and transported northwestward in the synoptic flow.

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Observations and modelling also demonstrate that during Flight 4, a LSC lake-breeze recirculation lofted pollutants from the CBL at the front, carried them in the southwesterly synoptic flow, they were caught in the downdraft over the lake, and then confined by the LSC onshore flow along the south shore of LSC. Modelling suggests that air parcels caught in the LSC recirculation travelled a helical trajectory along the LSC front toward Detroit/Windsor. Based upon measurements and model output, best estimates for recirculation times of the air parcels travelling in the LSC lake-breeze circulation range from 1.2 to 3.0 h.

This detailed, integrated approach led to the development of conceptual models that summarize the complex 3-D circulation patterns and its interaction with the synoptic flow. An attempt was then made at determining local SO₄²⁻ and OA production rates in the LSC lake-breeze circulation. Using the relationship between oxygenated organic aerosol (OOA)/ΔCO and SO₄²⁻/(SO₂ + SO₄²⁻) (Fig. 12) as a gauge of how much OOA is formed and a background formation rate for SO₄²⁻ ranging between 1–2% h⁻¹ taken from the model, the regional background formation rate for OOA was calculated to be 0.71–1.4 μg m⁻³ ppmv⁻¹ h⁻¹. Based on the interpretation of the LSC lake-breeze circulation, air masses in the CBL and behind the LSC front (AF) (Figs. 10a, 11) were compared to determine formation rates of OOA and SO₄²⁻. The maximum and minimum values (3.0 and 1.2 h, respectively) of the best estimates for recirculation time (or processing time) provide upper and lower bounds on the formation rates. The OOA formation rate, relative to excess CO, determined in the lake-breeze recirculation was found to be 2.5–6.2 μg m⁻³ ppmv⁻¹ h⁻¹ which shows an enhancement over the regional background formation values inferred above (0.71–1.4 μg m⁻³ ppmv⁻¹ h⁻¹). The SO₄²⁻ formation rate is estimated to range between 1.8–4.6% h⁻¹, also enhanced over the regional background formation rate. The enhanced formation rates relative to regional background rates implies that lake-breeze circulations are an important dynamic in the local, near-source formation of SO₄²⁻ and SOA. The presence of cumulus clouds associated with the lake-breeze fronts suggests that these enhancements could be due to cloud processes. Additionally, the effective confinement of pollutants along the

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Table 1. Summary of Twin Otter flights showing the synoptic wind direction (based on back-trajectories analyses), an air quality descriptor (AQ), the lake breeze type (LD = Low Deformation; MD = Medium Deformation; HD = High Deformation; details in Sills et al., 2011), and O_3 and SO_4^{2-} mean and maximum levels based on aircraft measurements.

Flt #	Date	Takeoff Time (LT)	Landing Time (LT)	Flight Duration (hh:mm)	Synoptic Wind Direction	AQ/Source Region	Lake Breeze Type	$[\text{O}_3]$ (ppbv)		$[\text{SO}_4^{2-}]$ ($\mu\text{g m}^{-3}$)	
								Mean	Max	Mean	max
1	23 June	08:52	11:15	02:23	Northwest	Good/Local sources	LD	32.3	39.2	0.14	0.33
2	23 June	13:30	15:57	02:27	Northwest	Good/Local sources	LD	39.9	57.7	0.34	1.89
3	23 June	17:43	18:16	00:33	Northwest	Good/Local sources	LD	40.0	47.4	0.39	0.74
4	25 June	11:01	13:24	02:23	Southwest	Poor/Local sources	LD	64.9	84.1	4.70	15.13
5	25 June	15:37	17:56	02:19	Southwest	Poor/Local sources	LD	73.0	105.5	6.92	29.07
6	25 June	19:13	20:07	00:54	Southwest	Poor/Local sources	LD	71.8	84.1	7.79	10.72
7	26 June	08:46	11:06	02:20	Southwest	Poor/Long range transport	MD	67.5	87.2	6.00	25.51
8	26 June	13:26	15:06	01:40	Southwest	Poor/Long range transport	MD	71.3	93.3	4.35	14.89
9	26 June	16:16	18:20	02:04	Southwest	Poor/Long range transport	MD	70.3	93.2	4.81	18.59
10	27 June	08:43	09:43	01:00	West	Detroit outflow	HD	43.3	65.8	2.18	9.73
11	27 June	11:17	13:41	02:24	West	Detroit outflow	HD	62.9	85.2	2.98	8.20
12	3 July	19:00	21:30	02:30	South then north	Poor/Long range transport	MD	66.2	102.6	2.32	11.78
13	7 July	04:35	07:17	02:42	Northwest	Detroit outflow	MD	45.4	79.0	0.56	1.87
14	7 July	13:44	16:20	02:36	Northwest	Detroit outflow	MD	63.6	87.2	0.75	3.53
15	8 July	11:36	14:14	02:38	West	Detroit outflow	HD	66.5	87.2	2.47	7.41
16	8 July	18:32	21:00	02:28	West	Detroit outflow	HD	69.8	81.0	2.36	4.65

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Table 2. Summary of trace gas and particle instrumentation on the Twin Otter aircraft.

Measurement	Principle of Operation	Instrument	Resolution (s)
Particle composition	aerosol mass spectrometry/ time of flight detection	Aerodyne C- AMS	30
O_3	UV absorption	TECO 49	5
CO	VUV resonance fluorescence	Aerolaser	1
SO_2	UV fluorescence	TECO 43S	10
NO/NO_2	chemiluminescence, photolysis	TECO 42S	20, alternating
Particle number	light scattering	TSI 7610 CNC	1
Particle size distribution (0.120–2 μm)	light scattering	PCASP (passive cavity absorption spectrometer probe)	1
Black carbon	light absorption	PSAP (particle soot absorption photometer)	1

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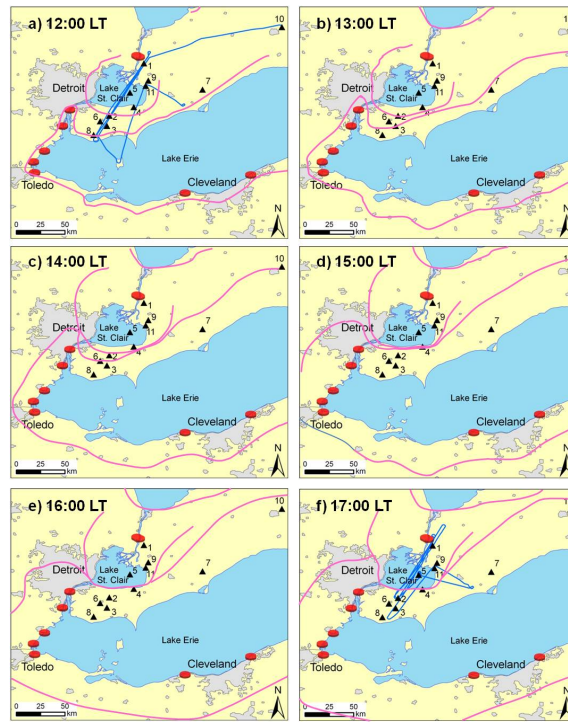


Fig. 1. Progression of lake-breeze fronts from 12:00–17:00 LT on 25 June. Position of lake-breeze fronts shown in magenta lines. Tracks for Flights 4 and 5 are shown as blue lines in (a) and (f) respectively. SO₂ major point sources shown as red circles (SO₂ > 10 000 tons yr⁻¹). Surface sites in black triangles: 1. Sombra, 2. Woodslee, 3. Cottam, 4. Lighthouse Cove, 5. LSC buoy, 6. Essex, 7. Ridgetown, 8. Harrow, 9. Bear Creek, 10. London, 11. Mitchell's Bay.

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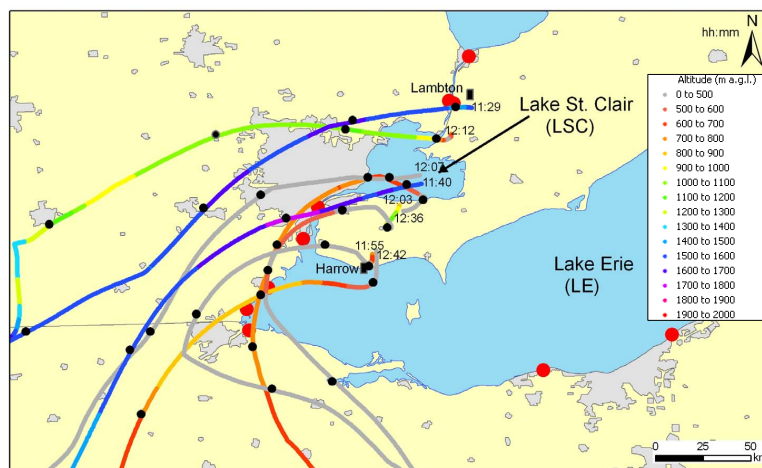


Fig. 2. High resolution back trajectories for 25 June arriving at selected locations and times along the Flight 4 aircraft track. Arrival times indicated as hh:mm. Trajectories are coloured by altitude (m a.g.l.). Black circles along trajectories every 4 h. SO₂ point sources shown as red circles (SO₂ > 10 000 tons yr⁻¹).

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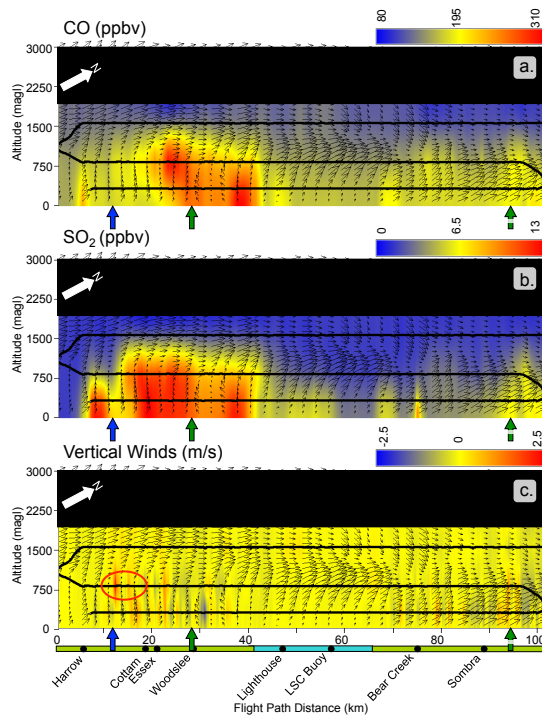


Fig. 5. Flight 4 cross-sections for (a) CO, (b) SO₂ and (c) vertical wind gusts along the axis of the transect. Horizontal wind direction overlaid (vertical wind component not included) and arrow size = wind speed magnitude. Land = green bar; lake = blue bar. Arrows indicate lake breeze fronts (blue = LE, green = LSC). Artefacts due to lack of data have been blacked out.

11537

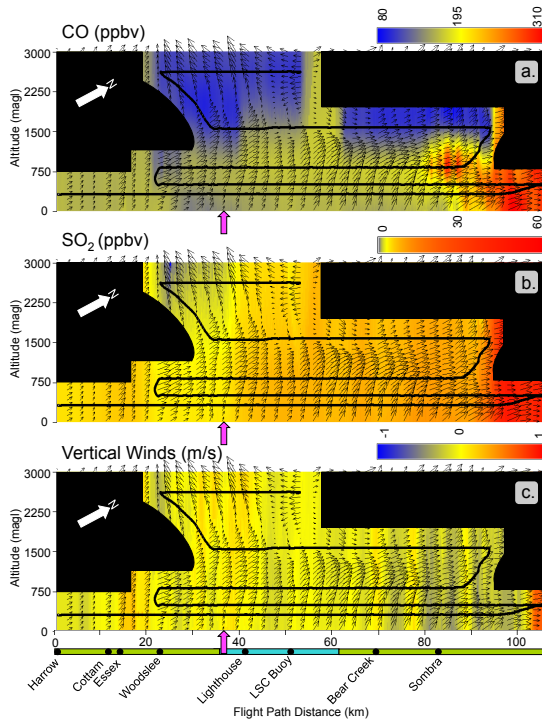


Fig. 6. Flight 5 cross-sections for CO, (b) SO₂ and (c) vertical wind gusts along the axis of the transect. Horizontal wind direction overlaid (vertical wind component not included) and arrow size = wind speed magnitude. Land = green bar; lake = blue bar. Magenta arrows indicate lake breeze fronts (magenta = LE/LSC merged front). Artefacts due to lack of data have been blacked out.

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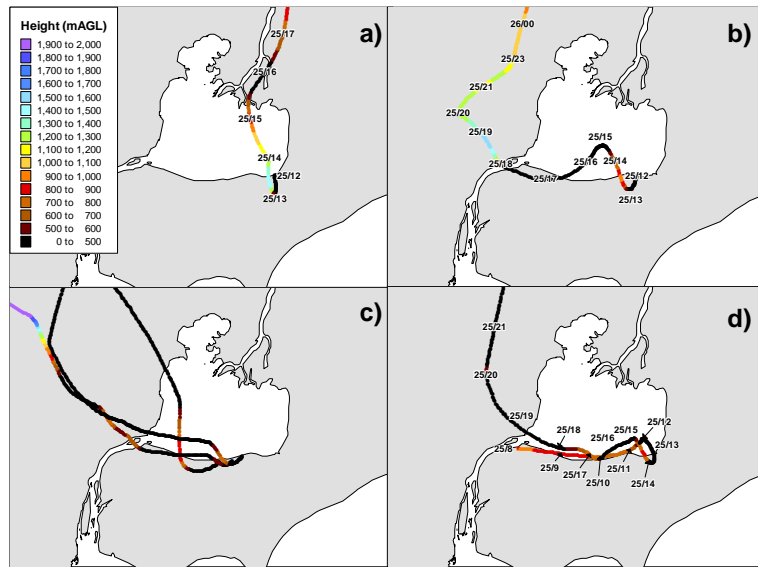


Fig. 9. Forward trajectories starting at (a) point 5; (b) point 4; (c) points 1, 2 and 3; along the LSC shoreline at 12:00 LT and (d) point 6 at 225 m a.g.l. above Windsor starting at 08:00 LT. Date/time (dd/hh) in panels (a), (b) and (d).

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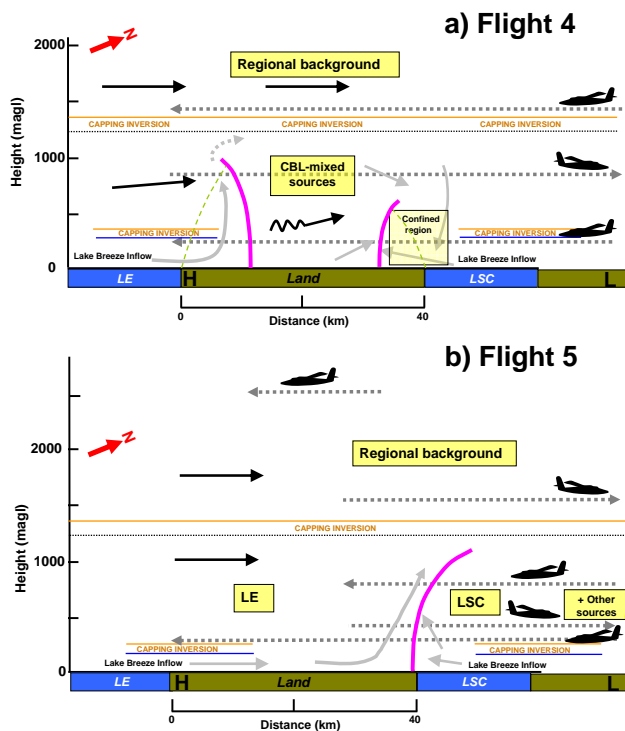


Fig. 10. Conceptual models of the lake breeze circulations for (a) Flight 4 and (b) Flight 5 based on the interpretation of aircraft, model and trajectory data. The black, squiggly arrow in the CBL depicts convective motion. H = Harrow, L = Lambton. The dashed grey arrow in Flight 4 illustrates a slower southwesterly flow due to effects from the LE return flow.

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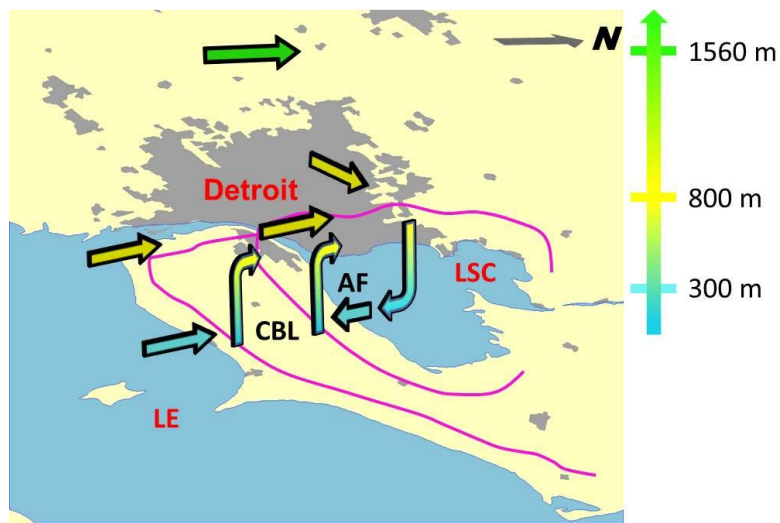


Fig. 11. Flight 4 perspective view. Surface position of lake-breeze fronts shown as magenta lines and the lake-breeze motions shown as arrows coloured according to altitude.

11543

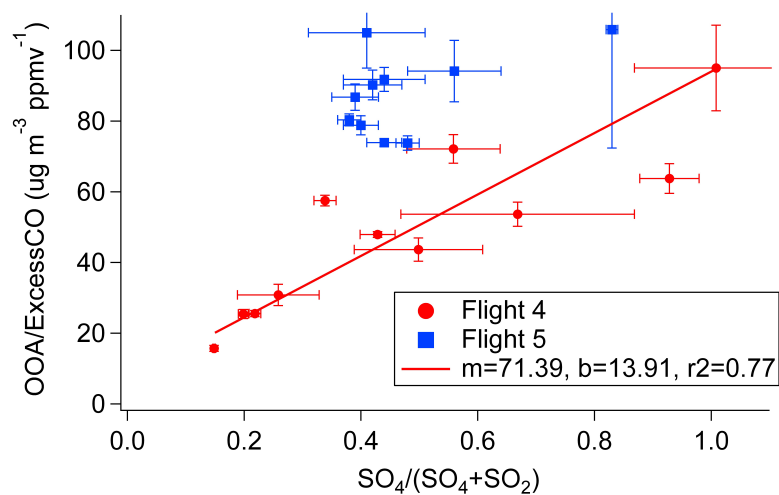


Fig. 12. Correlation of $OOA/\Delta CO$ with $SO_4^{2-}/(SO_2 + SO_4^{2-})$ across all air mass types for Flight 4 and 5.

11544

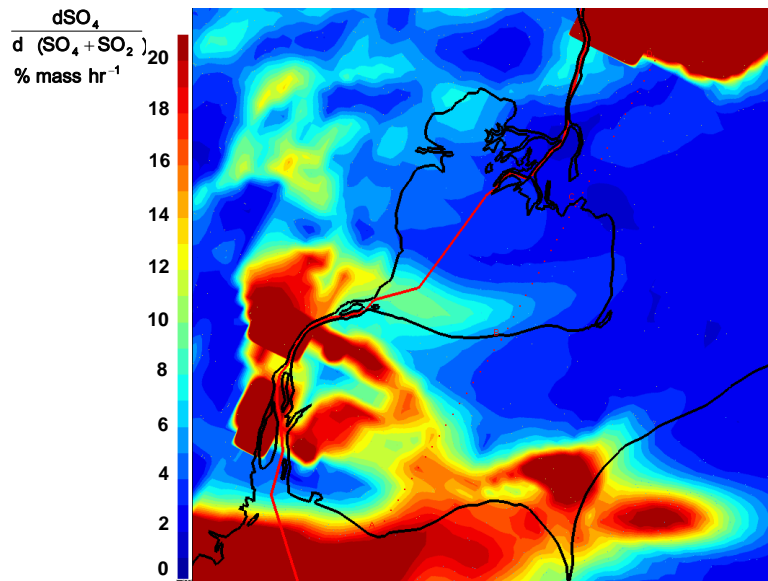


Fig. 13. Model-predicted SO_4^{2-} production rate in $\% \text{h}^{-1}$ at 12:00 LT for the study domain at 285 m a.g.l. Dashed red line is the aircraft transect.

11545

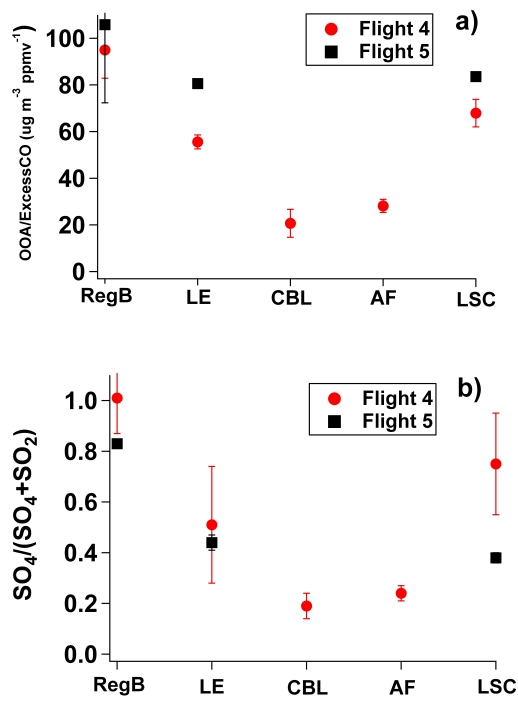


Fig. 14. (a) $\text{OOA}/\Delta\text{CO}$ and (b) $\text{SO}_4^{2-}/(\text{SO}_2 + \text{SO}_4^{2-})$ as a function of air mass type.

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